



Age and Growth of Marine and Translocated Freshwater Populations of *Atherina boyeri* Risso, 1810 (Atherinidae) in Turkey

Lale Gençoğlu¹, Şerife Gülsün Kirankaya^{1*} & F. Güler Ekmekçi²

¹Department of Biology, Faculty of Arts and Sciences, Duzce University, Konuralp Campus, 81620 Düzce, Turkey; E-mail: gkirankaya@gmail.com

²Department of Biology, Faculty Sciences, Hacettepe University, Beytepe Campus, 06800 Ankara, Turkey

Abstract: The sand smelt, *Atherina boyeri*, is a short-lived euryhaline fish that inhabits coastal and estuarine waters. It has also been illegally translocated into inland freshwater lakes and reservoirs for fisheries purposes in Turkey, and has become an invader in fresh water. The aim of this study was to compare the growth performance of translocated inland water populations and natural marine and brackish populations of *A. boyeri* in order to assess its adaptation potential to fresh water. The age, body length and weight parameters of native populations of *A. boyeri* were studied at six locations in the Mediterranean Sea, Black Sea, Sea of Marmara, Aegean Sea, and one in Köyceğiz Lake lagoon. The same parameters of translocated inland populations were studied in two reservoirs (Aslantaş and Hirfanlı) and in two freshwater lakes (Eğirdir and İznik). The findings show that *A. boyeri* exhibit similar growth performance in the marine, brackish, and fresh water studied, indicating that this species is successful in adaptation to various environmental conditions that favours its invasive potential in Turkey.

Key words: Sand smelt, invasive fish, inland populations, age, condition.

Introduction

The introduction and/or translocation of fish species is a common practice worldwide, especially since the last century. Fisheries and aquaculture are among the most significant causes of intentional fish transportation (LOCKWOOD et al. 2007, TARKAN et al. 2015). Although some introduced species provide new sources of income due to their commercial, recreational, and ornamental value, native and non-native species introductions frequently result in economic loss, both locally and globally. Today, the introduced and translocated species are considered to be among the most important human-mediated threats to the ecology and economy (CRIVELLI

1995, PIMENTEL et al. 2005, KETTUNEN et al. 2008, GOZLAN et al. 2010).

Turkey has the richest freshwater ichthyofauna in the Mediterranean region, with approximately 380 freshwater fish species (FREYHOF et al. 2014, ÇIÇEK et al. 2016, FROESE & PAULY 2019). Of these, 21 species are non-native and 28 native species are considered native translocated species (TARKAN et al. 2015). It has been reported that both non-native and translocated species can have adverse effects on native fish fauna due to their invasive characteristics (BOGUTSKAYA & NASEKA 2002); therefore, invasive species pose a serious threat to Turkey's rich freshwater fish biodiversity, in addition to such threats as pollution, excessive

*Corresponding author

water abstraction, and overfishing (EKMEKÇİ et al. 2013).

The sand smelt, *Atherina boyeri* Risso, 1810 (Atherinidae), is one of the most widespread translocated species in Turkey (EKMEKÇİ et al. 2013, TARKAN et al. 2015). *Atherina boyeri* is naturally a marine species widely distributed throughout the north-eastern coast of the Atlantic Ocean, Mediterranean Sea, Black Sea, and Caspian Sea (KOTTELAT & FREYHOF 2007). It is an euryhaline fish with a short life span that primarily inhabits coastal and estuarine waters. It has also been introduced into freshwater lakes and reservoirs for fishery purposes in many European countries, as well as in Turkey (KOTTELAT & FREYHOF 2007).

Atherina boyeri was first observed in the Turkish freshwater systems in Sapanca Lake (BATTALGİL 1941) and thereafter it has been observed in the Sakarya and Menderes river basins in the 1970s (ALTUN 1991). After the 1990s, the species has rapidly invaded the Turkish freshwater lakes and reservoirs, including the Aksu, Kızılırmak, Seyhan, Ceyhan, and Asi river basins (EKMEKÇİ et al. 2013). TARKAN et al. (2015) reported that *A. boyeri* has been deliberately introduced into the Turkish inland waters for fisheries, and that it has extensive self-sustaining populations in many natural and artificial lakes in Turkey. In fact, *A. boyeri* is currently one of the most valuable commercial (exported) fish species in the inland waters in Turkey (ATALAY et al. 2017).

Despite its commercial value, *A. boyeri* is considered an invasive and problematic fish species in Turkey (TARKAN et al. 2015), and its ecological and economic effects in inland waters are a cause for concern (KÜÇÜK et al. 2006, ATALAY et al. 2017). It is found out that the establishment success of *A. boyeri* in different ecosystems could be related to its growth, reproductive, and feeding characteristics (GENÇOĞLU & EKMEKÇİ 2016, GENÇOĞLU et al. 2017). However, there are a few studies on the species life-history traits and ecological effects in native marine habitats and recipient freshwater environments with the aim to assess its potential risks to inland waters. The fact that there are only a few life-history studies on the sand smelt populations in Turkey (GAYGUSUZ 2006, TARKAN et al. 2007, ÖZEREN 2009, GENÇOĞLU & EKMEKÇİ 2016, GENÇOĞLU et al. 2017) could be the reason why its risk ranking in freshwater systems has been considered to be moderate (TARKAN et al. 2015). Yet, a species categorisation of medium risk does not imply that it possesses no risk (TARKAN et al. 2015). For instance, GENÇOĞLU et al. (2017) report

that *A. boyeri* may have a negative effect on the zooplankton communities in Hirfanlı Reservoir in Turkey via selective predation.

In order to determine the effect of an invasive organism on the native species and on the ecosystem functioning, observational and experimental studies are required. Monitoring the life history characteristics of an invasive fish species in different locations is a component of such research (GARCIA-BERTHO 2007). To prevent the spread and establishment of new populations of *A. boyeri*, and to develop management strategies for its previously established populations, its biological characteristics should be determined in different habitats.

The objectives of the present study were to evaluate and compare the variability in growth parameters among six natural marine populations, one lagoon population, and four translocated freshwater populations of *A. boyeri*, in order to assess the species adaptation potential in freshwater ecosystems. To the best of our knowledge the present study is the first to concurrently examine the age and growth parameters of *A. boyeri* in a relatively wide range of ecologically different aquatic ecosystems.

Materials and Methods

The fish samples were collected between 2015 and 2017 from the Mediterranean Sea, Aegean Sea, Black Sea, Sea of Marmara, Köyceğiz Lake lagoon, and from multiple regions of Anatolia, including Aslantaş and Hirfanlı reservoirs, and the freshwater lakes Eğirdir and İznik (Fig. 1). The sampling was conducted at least three times at each sampling location during the spring months, when *A. boyeri* used the shores extensively. The samples were taken by a beach seine net.

The fish individuals caught were preserved in 4% formaldehyde solution and transported to the laboratory for further analysis. The total length (TL) was measured to the nearest 0.05 mm, and the weight (W) was measured to the nearest 0.001 g. The age determination was performed using scales and otoliths, and was validated using the Bhattacharya method and length frequency distribution. The scales were removed, cleansed with water, and placed on two slides, then evaluated twice by independent researchers using a binocular microscope (LAGLER 1966). The sagittal otoliths were cleansed with alcohol and water, and then evaluated twice by independent researchers under a microscope in 10% glycerin solution.

The length-weight relationship (LWR) was determined using linear regression analysis. Log-

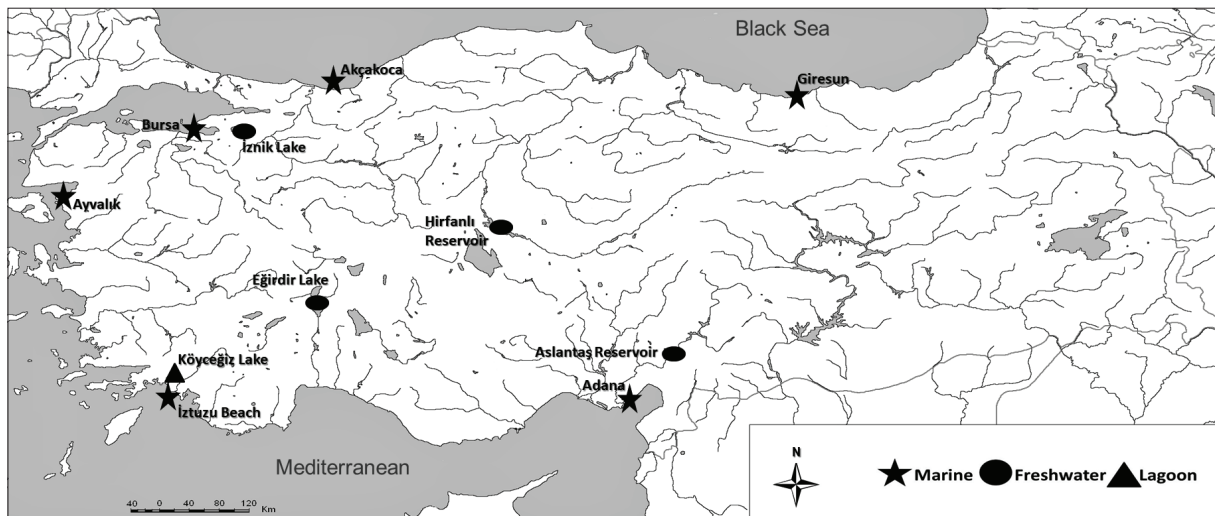


Fig. 1. Sampling locations of *Atherina boyeri*. The marine habitats are marked with a star, the freshwater habitats – with a circle, while the lagoon habitat – with a triangle.

transformed data were used to establish the LWR. The parameters a and b were calculated using least-squares regression (LE CREN 1951). Fulton's condition factor (K) was calculated by the equation $K=(W/L^3)\times 100$, where W is the total body weight (g) and L is the total length (cm) (LE CREN 1951). Both, the comparison of the slopes of the length-weight regression between sexes and the variation in coefficient b from 3 were tested using Student's t -test (ZAR 1999). Analysis of variance (ANOVA) followed by the Bonferroni post-hoc comparison test were performed using Microsoft Excel (2016) and Daniel's XL Toolbox add-in for Excel (version 7.3.2) to determine significance of the difference in length, weight, and condition factor values between populations and habitats.

The parameters of the von Bertalanffy growth equation (L_∞ – asymptotic length; k – growth coefficient; t_0 – hypothetical age at which length is 0) and the growth performance index [$\Phi' = 2\log(L_\infty) + \log(k)$] were calculated using the ELEFAN I method (GAYANILO et al. 1988). The physical and chemical parameters of the surface water at each sampling location, including the temperature (T), the electrical conductivity (C), and the salinity, were measured in situ using a YSI ProPlus Multiparameter water quality meter.

Results

The sampling locations had characteristic salinity values in accordance with the Venice Symposium Salinity System Specifications for marine, brackish water, and freshwater habitats (ANONYMOUS 1959). The salinity of marine environment was >15 ppt,

versus approximately 1.5 ppt of the lagoon, and <1 ppt of freshwater environment. The conductivity exhibited a pattern similar to that of salinity; the mean values ranged between 25,167 and 55,415 $\mu\text{S}/\text{cm}$ in the marine environment, from 235 to 1,551 $\mu\text{S}/\text{cm}$ in the freshwater environment, while the mean value in the lagoon was 3,096 $\mu\text{S}/\text{cm}$. During the sampling period, the mean water temperature ranged from 19.60–25.57°C, 24.60°C, and 14.75–22.40°C, in the marine, lagoon, and freshwater environment, respectively (Table 1).

In total, five age groups (0+–4+) were identified in the populations of *A. boyeri*. The dominant age group varied between sampling locations, while regarding habitats (marine, lagoon, and freshwater), the 1+ and 2+ age groups were dominant at all habitats (Fig. 2).

The total length range was from 21.25–142.96 mm in the marine populations, 27.93–97.03 mm in the lagoon population, and 23.87–115.65 mm in the freshwater populations. Except for the İztuzu (Aegean Sea) and Adana (eastern Mediterranean) populations, the individuals of >60 mm in length were dominant (Fig. 3). The ANOVA showed that there were significant differences in length frequencies between populations and habitats: between populations: $F=364.49$, $df=10$, and $P<0.05$; between marine and freshwater habitats: $F=6.05$, $df=1$, and $P<0.05$; freshwater – lagoon habitats: $F=73.66$, $df=1$, $P<0.05$; and marine – lagoon habitats: $F=57.72$, $df=1$, and $P<0.05$). The Bonferroni post-hoc comparison also showed that there were significant differences in the length frequencies between habitats.

The body weight range was 0.04–24.18 g in

Table 1. Sampling locations, number of *Atherina boyeri* individuals caught (N), and physical and chemical parameters measured: water temperature (T), conductivity (C), and salinity.

| Sampling locations | | N | T (°C) Mean ± SD | C (µS/cm) Mean ± SD | Salinity (ppt) Mean ± SD | |
|--------------------|---------------------------|---------------------------------------------------|---------------------|------------------------|-----------------------------|------------|
| Marine | Giresun (Black Sea) | N 40° 54' 55,98" E 38° 25' 0,24" | 485 | 20.50±5.23 | 26,282±2,158 | 16.15±2.52 |
| | Akçakoca (Black Sea) | N 41° 04' 39,5" E 31° 04' 05,2" | 373 | 19.60±0.99 | 25,167±950 | 17.55±0.16 |
| | Bursa (Sea of Marmara) | N 40° 22' 46,39" E 28° 52' 10,83" (Mudanya) | 291 | 19.90±4.92 | 34,725±10,981 | 23.40±5.48 |
| | | N 40° 47' 70,46" E 29° 08' 38,72" (Gemlik) | | | | |
| | Ayvalık (Aegean Sea) | N 39° 20' 36,17" E 26° 41' 18,66" | 378 | 22.33±3.44 | 55,415±11,431 | 39.13±8.43 |
| | İztuzu Beach (Aegean Sea) | N 36° 46' 49,1" E 28° 37' 55,3" | 275 | 25.57±9.13 | 38,752±18,333 | 24.75±6.39 |
| | Adana (Mediterranean Sea) | N 36° 44' 46,2" E 35° 38' 06,3" | 346 | 25.57±5.76 | 50,478±4,830 | 33.36±5.85 |
| Lagoon | Köyceğiz Lake (Muğla) | N 36° 57' 25,6" E 28° 40' 24,8" | 355 | 24.60±6.55 | 3,096±2,089 | 1.46±1.08 |
| Freshwater | İznik Lake | N 40° 22' 56,9" E 29° 34' 03,1" | 290 | 22.17±4.42 | 940±170 | 0.48±0.10 |
| | Eğirdir Lake | N 37° 52' 53,84" E 30° 49' 13,13" | 464 | 14.75±4.60 | 235±26 | 0.14 |
| | Hirfanlı Reservoir | N 39° 09' 44,04" E 33° 38' 14,96" | 369 | 19.93±4.12 | 1,551±320 | 0.83±0.15 |
| | Aslantaş Reservoir | N 37° 22' 57,2" E 36° 16' 38,2" | 409 | 22.40±9.48 | 370±10 | 0.19±0.04 |

the marine populations, 0.12–6.60 g in the lagoon population, and 0.08–10.94 g in the freshwater populations. The ANOVA showed that there were significant differences in weight frequencies between habitats: between populations: $F=363.62$, $df=10$, and $P<0.05$; between marine and freshwater habitats: $F=79.36$, $df=1$, and $P<0.05$; freshwater – lagoon: $F=69.19$, $df=1$, and $P<0.05$; and marine – lagoon habitats: $F=86.65$, $df=1$, and $P<0.05$). The Bonferroni post-hoc comparison also showed that there were significant differences in the weight frequencies between habitats.

The estimates of parameter b in the length-weight relationships varied between 2.77 and 3.35 in the populations, and the determination coefficient (r^2) values were close to 1. The difference in b values between the marine and freshwater populations was not significant (Student's t-test, $P>0.05$). The mean value of the condition factor ranged from 0.53 to

0.63 in the marine populations, from 0.53 to 0.61 in the freshwater populations, and it was 0.52 in the lagoon population (Table 2). There were significant differences in the condition factor between populations and habitats (ANOVA): between populations: $F=51.61$, $df=10$, and $P<0.05$; between marine and freshwater habitats: $F=627.66$, $df=1$, and $P<0.05$; freshwater – lagoon: $F=2400.13$, $df=1$, and $P<0.05$; and marine – lagoon habitats: $F=4890.16$, $df=1$, and $P<0.05$). The Bonferroni post-hoc comparison also showed that there were significant differences in the condition of *A. boyeri* between habitats. The growth coefficient (k) range was from 0.16–0.79 in the marine populations, 0.23–0.49 in the freshwater populations, and 0.29 in the lagoon population. The growth performance index values ranged from 7.67 to 9.96 in the marine populations, from 8.14 to 8.89 in the freshwater populations, and it was 8.10 in the lagoon population (Table 2).

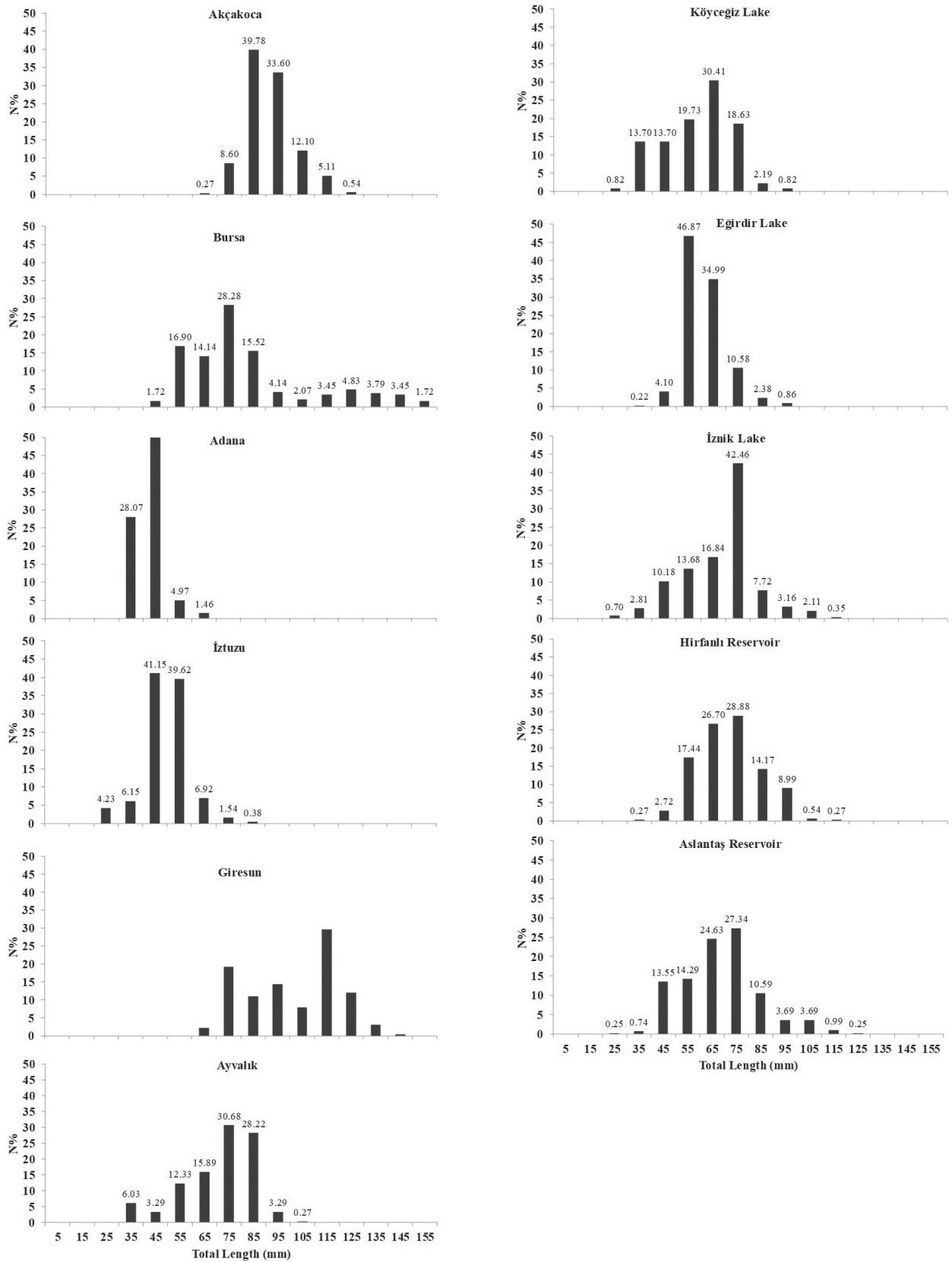


Fig. 2. Age distribution of *Atherina boyeri* at the studied locations. N – number of individuals (percentages, %); Age (years).

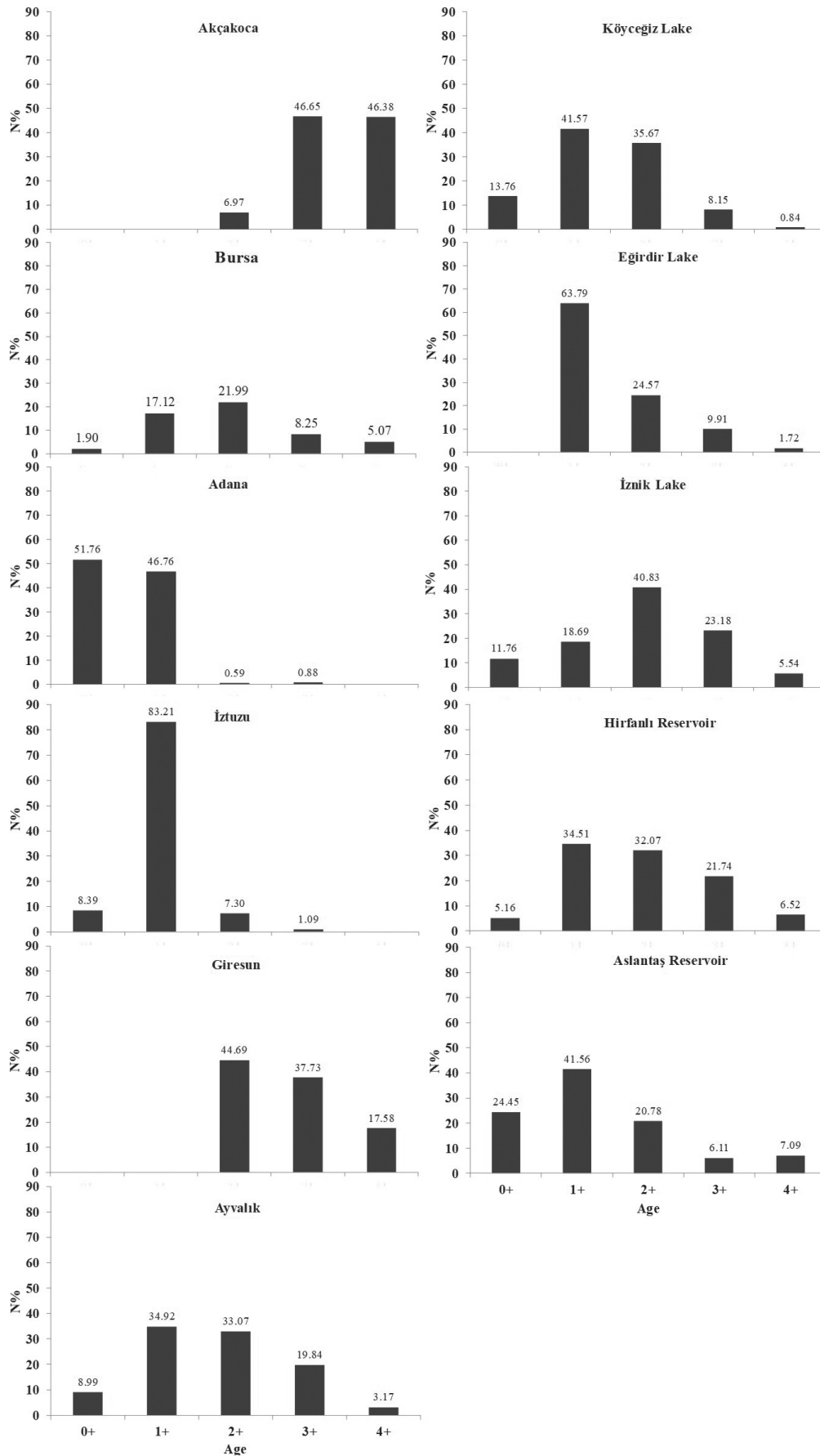


Fig. 3. Length frequency histograms of *Atherina boyeri*. N – number of individuals (percentages, %); Total length (mm).

Table 2. Estimated length-weight relationship and von Bertalanffy parameters, as well as growth performance indices of *Atherina boyeri* populations at the studied locations. TL – total length; W – total body weight; a and b – parameters of the equations; r^2 – coefficient of determination; * – exponent b differed significantly from 3; L_{∞} , k and t_0 – parameters of the von Bertalanffy equation. For further explanation, please see Materials and Methods.

| Sampling locations | TL (mm) | | W (g) | | Regression parameters | | | | Condition factor (Mean±SD) | The von Bertalanffy growth equation parameters | | | Growth performance index | |
|--------------------|--------------------|-------|--------|------|-----------------------|--------|-------|---------|----------------------------|------------------------------------------------|------|-------|--------------------------|------|
| | Min | Max | Min | Max | a | b | r^2 | t-value | | L_{∞} (cm) | k | t_0 | | |
| Marine | Giresun | 54.45 | 142.96 | 1.13 | 17.24 | -1.904 | 2.767 | 0.957 | 10.13* | 0.59±0.07 | 14.3 | 0.36 | -0.20 | 8.90 |
| | Akçakoca | 69.77 | 129.77 | 1.85 | 17.87 | -2.351 | 3.271 | 0.905 | 4.297* | 0.63±0.07 | 13.6 | 0.55 | -0.57 | 9.23 |
| | Bursa | 43.18 | 139.24 | 0.52 | 24.18 | -1.953 | 2.776 | 0.992 | 20.36* | 0.62±0.08 | 16.4 | 0.79 | -0.84 | 9.96 |
| Lagoon | Ayvalık | 33.36 | 102.43 | 0.24 | 5.90 | -2.293 | 3.119 | 0.980 | 5.17* | 0.56±0.07 | 12.8 | 0.16 | -0.34 | 7.87 |
| | Iztuzu | 21.25 | 83.71 | 0.04 | 3.80 | -2.454 | 3.353 | 0.978 | 11.39* | 0.53±0.07 | 9.9 | 0.42 | -0.82 | 8.33 |
| | Adana | 33.48 | 69.13 | 0.20 | 2.01 | -2.275 | 3.075 | 0.929 | 1.63 | 0.53±0.06 | 9.6 | 0.23 | -0.91 | 7.67 |
| Freshwater | Köyceğiz Lake | 27.93 | 97.03 | 0.12 | 6.60 | -2.355 | 3.206 | 0.953 | 5.57* | 0.52±0.09 | 10.6 | 0.29 | -0.66 | 8.10 |
| | Iznik Lake | 27.21 | 115.65 | 0.08 | 9.79 | -2.279 | 3.131 | 0.9658 | 5.17* | 0.59±0.09 | 12.1 | 0.49 | -0.95 | 8.89 |
| | Eğirdir Lake | 34.43 | 95.45 | 0.39 | 6.67 | -2.233 | 3.075 | 0.941 | 1.965* | 0.53±0.10 | 10.9 | 0.29 | -0.62 | 8.14 |
| Freshwater | Hirfanlı Reservoir | 38.64 | 113.20 | 0.34 | 10.94 | -2.160 | 3.011 | 0.917 | 0.23 | 0.56±0.07 | 14.1 | 0.23 | -0.27 | 8.43 |
| | Aslantaş Reservoir | 23.87 | 115.35 | 0.11 | 9.64 | -2.247 | 3.106 | 0.978 | 4.61* | 0.61±0.07 | 16.0 | 0.23 | -0.37 | 8.68 |

Discussion

Populations of *A. boyeri* in marine, brackish, and fresh water are exposed to several environmental stressors and such widely-fluctuating conditions as salinity, conductivity, and temperature. Although salinity and temperature are thought to be the primary environmental constraints that directly or indirectly affect fish communities (WOOTTON 1994), in the present study the factors with the greatest effects on the growth performance of *A. boyeri* are salinity and conductivity because of the compulsory osmoregulation adaptation.

Except for the Black Sea populations, the age composition of the marine, brackish and freshwater populations of *A. boyeri* does not differ significantly. The length-frequency distribution of *A. boyeri* follows a similar pattern as the age composition; larger individuals are predominant in the Black Sea populations, whereas smaller individuals dominate in other populations. Our findings and those reported in previous studies indicate that the life span of *A. boyeri* covers a short period of up to 4+ years (ALTUN 1986, CREECH 1992, ROSECCHI & CRIVELLI 1992, BOZDAĞ 1999, LEONARDOS & SINIS 2000, ANDREU-SOLER et al. 2003, BARTULOVIC et al. 2004a, KOUTRAKIS et al. 2004, POMBOS et al. 2005, SEZEN 2005, GAYGUSUZ 2006, KÜÇÜK et al. 2006, TARKAN et al. 2007, ÖZEREN 2009, PATIMAR et al. 2009, KESKIN & GAYGUSUZ 2010, BÖK et al. 2011, GÜRKAN et al. 2014, SEVINÇ 2014, APAYDIN YAĞCI et al. 2015, İLHAN & SARI 2015, GENÇOĞLU & EKMEKÇI 2016). In addition to the short life span, *A. boyeri* is also known to have early maturation and the ability to spawn repeatedly (CREECH 1992, ROSECCHI & CRIVELLI 1992, PATIMAR et al. 2009, GENÇOĞLU & EKMEKÇI 2016). Early maturation and multiple spawning are typical characteristics of the invasive species, along with a short life span (BOGUTSKAYA & NASEKA 2002, GRABOWSKA & PRZYBYLSKI 2014).

The exponent *b* in the length-weight relationship in the present study varies between 2 and 4, but often is a value close to 3; a value of 3 indicates isometric growth and values other than 3 indicate allometric growth (TESCH 1971). The present findings show that positive allometric growth ($b>3$, $P<0.05$) is common in the marine, lagoon, and freshwater populations of *A. boyeri* in Turkey, whereas the growth model is negative allometric in two marine populations (Giresun and Bursa: $b<3$ and $P<0.05$), and isometric in two freshwater populations (Adana and Hirfanlı: $b=3$ and $P>0.05$). In other words, the weight increase rates are slightly lower than the length increase in the Giresun, Bursa, Adana, and Hirfanlı populations,

although there is a similar growth rate in other native and translocated populations. The LWR can be indicative of spatial and temporal variations related to water temperature, food availability, and reproductive activity (WOOTTON 1994). The LWR parameters a and b are affected by a number of factors, including sex, gonad maturity, health status, season, habitat, nutrition, environmental conditions as temperature and salinity, stomach fullness, general fish condition, differences in length range of fish individuals, and collection gear (TESCH 1971, FROESE 2006). According to the present findings, growth performance of *A. boyeri* is not affected by the environmental conditions, especially, salinity in the freshwater habitats, in contrast to our expectations.

Studies on population dynamics have shown that high values of the condition factor indicate favorable environmental conditions, including habitat and prey availability (BLACKWELL et al. 2000). The condition factor of *A. boyeri* in different habitats with varying salinity shows similar values in the present study. Moreover, condition factor values in different habitats reported previously (BOZDAĞ 1999, GAYGUSUZ 2006, ÖZEREN 2009, GENÇOĞLU & EKMEKÇİ 2016) are similar to our findings. Differences in the condition factor of each population in the present study could be considered indicative of the feeding conditions. *Atherina boyeri* is primarily a zooplanktivorous fish species with opportunistic feeding behaviour, which prefers the most abundant and large organisms in different ecosystems (GON & BEN-TUVIA 1983, MANTILACCI et al. 1990, DANILOVA 1991, ROSECCHI & CRIVELLI 1992, TRABELSI et al. 1994, BARTULOVIC et al. 2004b, CHRISAFI et al. 2007, DOULKA et al. 2012, APAYDIN YAĞCI et al. 2013). For example, in Hirfanlı Reservoir, a freshwater sampling location in the present study, *A. boyeri* prefers large food organisms, such as cladocerans, copepods, and even its own larvae, in accordance with variations in seasonal abundance of these organisms, and avoids small rotiferans even though the rotiferans are abundant throughout the year (YİĞİT & ALTINDAĞ 2005, GENÇOĞLU et al. 2017).

Other studies on the feeding behaviour of *A. boyeri* have shown that when the zooplankton is scarce due to a season or habitat, this species tends to feed on benthic organisms depending on their abundance (ROSECCHI & CRIVELLI 1992, TRABELSI et al. 1994, BARTULOVIC et al. 2004b). Due to this favoured feeding behaviour, it could be expected that *A. boyeri* consumes the most available and beneficial food organisms with the lowest energy cost; therefore, feeding conditions that could be challenging and unfavorable for growth performance

of many other fish species do not affect the growth performance of *A. boyeri* in different habitats.

The von Bertalanffy equation parameters and growth performance index values for *A. boyeri* in marine environments have not been calculated in earlier studies (BOZDAĞ 1999, KESKIN & GAYGUSUZ 2010, BÖK et al. 2011, GÜRKAN et al. 2014) and, therefore, comparisons with the present study cannot be made. However, our findings are similar to those of previous studies on populations in lagoons (CREECH 1992, ROSECCHI & CRIVELLI 1992, LEONARDOS & SINIS 2000, ANDREU-SOLER et al. 2003, BARTULOVIC et al. 2004a, KOUTRAKIS et al. 2004, POMBO et al. 2005, SEZEN 2005, PATIMAR et al. 2009) and fresh water (ALTUN 1986, GAYGUSUZ 2006, KÜÇÜK et al. 2006, TARKAN et al. 2007, ÖZEREN 2009, SEVINÇ 2014, APAYDIN YAĞCI et al. 2015, İLHAN & SARI 2015, GENÇOĞLU & EKMEKÇİ 2016). The von Bertalanffy equation parameters for *A. boyeri* populations in both the present and previous studies indicate that these parameters are not significantly affected by habitat differences.

The translocation of any given species to a new environment can induce changes in its life-history traits, especially if an increase in energy cost (e.g., for physiological regulation, survival, escaping new predators and foraging) is involved (ROSECCHI et al. 2001). In the present study, the growth parameters, including condition factor, LWR, and von Bertalanffy growth parameters, at 11 sampling locations, are in general similar despite the significant differences in salinity and conductivity. The observed similarity in growth performance might have been due to the physiological tolerance and adaptation success of *A. boyeri*. This species is known to live along the coast of the seas surrounding Anatolia and in its lagoon systems. HENGEVELD & VAN DEN BOSCH (1996) suggest that the rate of spread is higher when a species has been introduced to an environment close to its native range. This may explain the rapid expansion and dispersal of *A. boyeri* as the lower parts of rivers connected with the seas or the lagoon provide suitable habitats for *A. boyeri*. Many atherinids, especially *A. boyeri*, have a high degree of intraspecific morphological variability that can be linked to their native estuarine and lagoon habitats. Such habitats are physically highly variable which has resulted in the atherinid morphology, physiology, and behaviour adapting to a wide range of conditions. This plasticity pre-adapts the atherinids to invade fresh waters (BAMBER & HENDERSON 1988). The present findings suggest that the invasive success of *A. boyeri* is facilitated by its ability to tolerate salinity changes via physiological plasticity. Despite

of the salinity changes no significant variation in the growth performance of this species has been observed. Further studies are required to determine more clearly the role of the physiological tolerance of *A. boyeri* for its invasive success and potential effects on the ecosystems.

Acknowledgements: This study was funded by the Scientific and Technological Research Council of Turkey (TÜBİTAK) with Project Number: TOVAG-114O809. We thank Dr Baran Yoğurtçuoğlu for his help with the sampling and collection of material. Also we would like to thank the anonymous reviewers for their constructive comments and suggestions on the manuscript.

References

- ALTUN Ö. 1986. The biology and ontogenetic development of the boyer's sand smelt (*Atherina (hepsetia) boyeri* Risso 1810) in Küçükçekmece Lake. PhD Thesis, İstanbul University, Institute of Science, İstanbul, 47 p. (In Turkish, English Abstract)
- ALTUN Ö. 1991. Morphology of the sand smelt (*Atherina boyeri* Risso, 1810) in Lake Küçükçekmece. Doğa-Turkish Journal of Zoology 15: 64-75. (In Turkish, English Abstract)
- ANDREU-SOLER A., OLIVA-PATERNA F. J., FERNÁNDEZ-DELGADO C. & TORRALVA M. 2003. Age and growth of the sand smelt, *Atherina boyeri* (Risso, 1810), in the Mar Menor coastal lagoon (SE Iberian Peninsula). Journal of Applied Ichthyology 19 (4): 202–208.
- ANONYMOUS 1959. Symposium on the classification of brackish waters. Venice, 8–14th April 1958. Archivio di Oceanografia e Limnologia, Volume 11. Supplemento.
- APAYDIN YAĞCI M., YAĞCI A., BILGIN F., ATAY R., DÖLCÜ B., UYSAL R., CESUR M., BOSTAN H. & YEĞEN V. 2013. The effects of introduced sand smelt (*Atherina boyeri* Risso, 1810) on food chain in Eğirdir Lake. TAGEM/HAYSÜD Project Report (Project Number: 2010/09/01/01), Isparta, 332 p. (In Turkish, English Abstract)
- APAYDIN YAĞCI M., ALP A., YAĞCI A. & CESUR M. 2015. Growth and reproduction of sandsmelt *Atherina boyeri* Risso, 1810 in Lake Eğirdir, Isparta, Turkey. Indian Journal of Fisheries 62 (1): 1–5.
- ATALAY M. A., KIRANKAYA Ş. G. & EKMEKÇİ F. G. 2017. The current status of gibel carp and sand smelt in Turkey's inland fisheries. Yunus Research Bulletin 17 (1): 41–57.
- BAMBER R. N. & HENDERSON P. A. 1988. Pre-adaptive plasticity in atherinids and the estuarine seat of teleost evolution. Journal of Fish Biology 33 (Supplement A): 17–23.
- BARTULOVIĆ V., GLAMUZINA B., CONIDES A., DULČIĆ J., LUČIĆ D., NJIRE J. & KOŽUL V. 2004a. Age, growth, mortality and sex ratio of sand smelt, *Atherina boyeri* Risso, 1810 (Pisces: Atherinidae) in the estuary of the Mala Neretva River (middle-eastern Adriatic, Croatia). Journal of Applied Ichthyology 20 (5): 427–430.
- BARTULOVIĆ V., LUČIĆ D., CONIDES A. J., GLAMUZINA B., DULČIĆ J., HAFNER D. & BATISTIĆ M. 2004b. Food of sand smelt, *Atherina boyeri* Risso, 1810 (Pisces: Atherinidae) in the estuary of the Mala Neretva River (middle-eastern Adriatic, Croatia). Scientia Marina 68 (4): 597–603.
- BATTALGİL F. 1941. Les poissons des Eaux Douces de la Turquie. Journal of İstanbul University, Faculty of Science [İstanbul Üniversitesi Fen Fakültesi Mecmuası], Seri B, 6: 170–186.
- BLACKWELL B. G., BROWN M. L. & WILLIS D. W. 2000. Relative weight (Wr) status and current use in fisheries assessment and management. Reviews in Fisheries Science 8 (1): 1–44.
- BOGUTSKAYA N. G. & NASEKA A. M. 2002. An overview of nonindigenous fishes in inland waters of Russia. Proceedings of the Zoological Institute, Russian Academy of Sciences, 296: 21–30.
- BOZDAĞ G. 1999. Researches about biology of the silverside (*Atherina boyeri* Risso, 1810) distributing in İzmir Bay. MSc Thesis, Ege University, Institute of Science, İzmir, Turkey, 44 p. (In Turkish, English Abstract)
- BÖK T. D., GÖKTÜRK D., KAHRAMAN A. E. & ALIÇLI T. Z. 2011. Length-weight relationships of 34 fish species from the Sea of Marmara, Turkey. Journal of Animal and Veterinary Advances 10 (23): 3037–3042.
- CHRISAFI E., KASPIRIS P. & KATSELIS G. 2007. Feeding habits of sand smelt (*Atherina boyeri*, Risso 1810) in Trichonis Lake (Western Greece). Journal of Applied Ichthyology 23 (3): 209–214.
- CREECH S. 1992. A study of population biology of *Atherina boyeri* Risso, 1810 in Aberthaw Lagoon, on the Bristol Channel, in South Wales. Journal of Fish Biology 41 (2): 277–286.
- CRIVELLI A. J. 1995. Are fish introductions a threat to endemic freshwater fishes in the Northern Mediterranean Region? Biological Conservation 72 (2): 311–319.
- ÇİÇEK E., SUNGUR BİRECİKLİĞİL S. & FRICKE R. 2016. Addenda and errata of: Freshwater fishes of Turkey: a revised and updated annotated checklist. FishTaxa 1 (2): 116–117.
- DANILOVA M. M. 1991. Diet of juvenile silversides, *Atherina boyeri*, from the Black Sea. Journal of Ichthyology 31 (1): 137–145.
- DOULKA E., KEHAYIAS G., CHALKIA E. & LEONARDOS I. D. 2012. Feeding strategies of *Atherina boyeri* (Risso 1810) in a freshwater ecosystem. Journal of Applied Ichthyology 29 (1): 200–207.
- EKMEKÇİ F. G., KIRANKAYA Ş. G., GENÇOĞLU L. & YOĞURTÇUOĞLU B. 2013. Present status of invasive fishes in inland waters of Turkey and assessment of the effects of invasion. İstanbul University Journal of Fisheries and Aquatic Sciences [İstanbul Üniversitesi Su Ürünleri Dergisi] 28 (1): 105–140. (In Turkish, English Abstract)
- FREYHOF J., EKMEKÇİ F. G., ALİ A., KHAMEES A. R., ÖZULUĞ M., HAMİDAN N., KÜÇÜK F. & SMITH K. G. 2014. Chapter 3. Freshwater fishes. Pp. 19–42. In: Smith K. G., Barrios V., Darwall W. R. T. & Numa C. (Eds.) The Status and Distribution of Freshwater Biodiversity in the Eastern Mediterranean. Cambridge, UK, Malaga, Spain, and Gland, Switzerland: IUCN, xvi, 129 p.
- FROESE R. 2006. Cube law, condition factor and weight-length relationships: metaanalysis and recommendations. Journal of Applied Ichthyology 22 (4): 241–253.
- FROESE R. & PAULY D. (Eds.) 2019. FishBase. World Wide Web Electronic Publication, www.fishbase.org, Version (08/2019).
- GARCIA-BERTHOU E. 2007. The characteristics of invasive fishes: what has been learned so far? Journal of Fish Biology 71 (Supplement sd): 33–55.
- GAYANILO F. C., JR., SORIANO M. & PAULY D. 1988. A draft guide to the Compleat ELEFAN. ICLARM Software 2. Manila, Philippines: International Center for Living Aquatic Resources Management, 65 p.
- GAYGUSUZ Ö. 2006. Reproduction and growth biology of the

- sandsmelt (*Atherina boyeri* Risso, 1810) living in the İznik Lake. MSc Thesis, İstanbul University, Institute of Science, İstanbul, Turkey, 56 p. (In Turkish, English Abstract)
- GENÇOĞLU L. & EKMEKÇİ F. G. 2016. Growth and reproduction of a marine fish, *Atherina boyeri*, Risso 1810 in a freshwater ecosystem. *Turkish Journal of Zoology*, 40: 534–542.
- GENÇOĞLU L., KIRANKAYA Ş. G., YOĞURTÇUOĞLU B. & EKMEKÇİ F. G. 2017. Feeding properties of the translocated marine fish sand smelt *Atherina boyeri* Risso, 1810 (Atherinidae) in a freshwater reservoir. *Acta Zoologica Bulgarica*, Supplement 9: 131–138.
- GON O. & BEN-TUVIA A. 1983. The biology of Boyer's sand smelt, *Atherina boyeri* Risso in the Bardawil Lagoon on the Mediterranean coasts of Sinai. *Journal of Fish Biology*, 22 (5): 537–547.
- GOZLAN R. E., BRITTON J. R., COWX I. & COPP G. H. 2010. Current knowledge on non-native freshwater fish introductions. *Journal of Fish Biology* 76 (4): 751–786.
- GRABOWSKA, J. & PRZYBYLSKI M. 2014. Life-history traits of non-native freshwater fish invaders differentiate them from natives in the Central European bioregion. *Reviews in Fish Biology and Fisheries*, DOI: 10.1007/s11160-014-9375-5.
- GÜRKAN E., BAYHAN Ş., AKÇINAR B. & TAŞKAVAK S. C. 2014. Length-weight relationship of fish from shallow waters of Candarli Bay (North Aegean Sea, Turkey). *Pakistan Journal of Zoology* 42 (4): 495–498.
- HENGEVELD R. & VAN DEN BOSCH F. 1996. Predicting the rate of spread of introduced animals and plants. *Wildlife Biology* 2 (3): 151–158.
- İLHAN A. & SARI H. M. 2015. Length-weight relationships of fish species in Marmara Lake, West Anatolia, Turkey. *Croatian Journal of Fisheries* 73 (1): 30–32.
- KESKIN C. & GAYGUSUZ O. 2010. Length-weight relationships of fishes in shallow waters of Erdek Bay (Sea of Marmara, Turkey). *IUFS Journal of Biology* 69 (2): 87–94.
- KETTUNEN M., GENOVEAI P., GOLLASCH S., PAGAD S., STARFINGER U., BRINK P. & SHINE, C. 2008. Technical support to EU strategy on invasive species (IS). Assessment of the impacts of IS in Europe and the EU. Final Module Report for the European Commission, Institute for European Environmental Policy (IEPP), Brussels, Belgium, 40 p. +Annexes.
- KOTTELAT M. & FREYHOF J. 2007. *Handbook of European freshwater fishes*. Berlin, Germany: Kottelat, Cornol, Switzerland and Freyhof, 646 p.
- KOUTRAKIS E. T., KAMIDIS N. I. & LEONARDOS I. D. 2004. Age, growth and mortality of a semi-isolated lagoon population of sand smelt, *Atherina boyeri* (Risso, 1810) (Pisces: Atherinidae) in an estuarine system of northern Greece. *Journal of Applied Ichthyology* 20 (5): 382–388.
- KÜÇÜK F., GÜLLE İ., GÜÇLÜ S. S., GÜMÜŞ E. & DEMİR O. 2006. Effect on fishery and lake ecosystem of non-native sand smelt (*Atherina boyeri* Risso, 1810) in Eğirdir Lake. 1st National Fisheries and Reservoir Management Symposium [1. Ulusal Balıklandırma ve Rezervuar Yönetimi Sempozyumu], 7–9 February 2006, Antalya, Turkey. (In Turkish, English Abstract)
- LAGLER K. F. 1966. *Freshwater fishery biology*. Dubuque, Iowa: Wm. C. Brown Company Publishers, 421 p.
- LE CREN E. D. 1951. The length-weight relationship and seasonal cycle in gonad weight and condition in the perch (*Perca fluviatilis*). *Journal of Animal Ecology* 20 (2): 201–219.
- LEONARDOS I. & SINIS A. 2000. Age, growth and mortality of *Atherina boyeri* Risso, 1810 (Pisces: Atherinidae) in the Mesolongi and Elotikon lagoons (W. Greece). *Fisheries Research* 45: 81–91.
- LOCKWOOD J. L., HOOPES M. F. & MARCHETTI M. P. 2007. *Invasion Ecology*. Wiley-Blackwell Publishing, 304 p.
- MANTILACCI L., MEARELLI M., GIOVITNAZZO G. & LORENZON M. 1990. Growth and nourishment of the sand smelt (*Atherina boyeri* Risso) of Lake Trasimeno. *Rivista di Idrobiologia* 29: 309–327. (In Italian, English Summary)
- ÖZEREN S. C. 2009. Age, growth and reproductive biology of the sand smelt *Atherina boyeri*, Risso 1810 (Pisces: Atherinidae) in Lake İznik, Turkey. *Journal of Fisheries International* 4: 34–39.
- PATIMAR R., YOUSEFI M. & HOSIENI S. M. 2009. Age, growth and reproduction of the sand smelt *Atherina boyeri* Risso, 1810 in the Gomishan wetland – southeast Caspian Sea. *Estuarine, Coastal and Shelf Science* 81 (4): 457–462.
- PIMENTEL D., ZUNIGA R. & MORRISON D. 2005. Update on the environmental and economic costs associated with alien-invasive species in the United States. *Ecological Economics* 52 (3): 273–288.
- POMBO L., ELLIOTT M. & REBELO J. E. 2005. Ecology, age and growth of *Atherina boyeri* and *Atherina presbyter* in the Ria de Aveiro, Portugal. *Cybum* 29 (1): 47–55.
- ROSECCI E. & CRIVELLI A. J. 1992. Study of a sand smelt (*Atherina boyeri* Risso, 1810) population reproducing in fresh water. *Ecology of Freshwater Fish*, 1: 77–85.
- ROSECCI E., THOMAS F. & CRIVELLI A. J. 2001. Can life-history traits predict the fate of introduced species? A case study on two cyprinid fish in southern France. *Freshwater Biology* 46 (6): 845–853.
- SEVINÇ N. 2014. Some biological characteristics of sand smelt (*Atherina boyeri* Risso, 1810) in Lake İznik (Bursa-Turkey). MSc Thesis, Marmara University, Institute of Science, İstanbul, Turkey, 53 p. (In Turkish, English Abstract)
- SEZEN B. 2005. Research on characteristics of sand smelt (*Atherina boyeri* Risso, 1810) population in homa lagoon (İzmir). MSc Thesis, Ege University, Institute of Science, İzmir, Turkey, 85 p. (In Turkish, English Abstract)
- TARKAN S., BİLGE G., SEZEN B., TARKAN A., GAYGUSUZ O., GÜRSOY C., FİLİZ H. & ACIPINAR H. 2007. Variations in growth and life history traits of sand smelt, *Atherina boyeri*, populations from different water bodies of Turkey: Influence of environmental factors. *Rapport Commission Internationale pour l'exploration scientifique de la Mer Méditerranée* (CIESM): 38: 611.
- TARKAN A. S., MARR S. M. & EKMEKÇİ F. G. 2015. Non-native and translocated freshwater fish species in Turkey. *FiSHMED Fishes in Mediterranean Environments* 003: 28 p.
- TESCH F. W. 1971. Age and growth. In: Ricker W. E. (Ed.), *Methods for assessment of fish production in fresh waters*. Oxford: Blackwell Scientific Publications, pp. 99–130.
- TRABELSI M., KARTAS F. & QUIGNARD J. P. 1994. Comparison of diet between a marine and a lagoonal of *Atherina boyeri* from Tunisian coasts. *Vie et Milieu* 44 (2): 117–123.
- WOOTTON R. J. 1994. *Ecology of Teleost Fishes*, London: Chapman & Hall, 404 p.
- YİĞİT S. & ALTINDAĞ A. 2005. A taxonomical study on the zooplankton fauna of Hirfanlı Dam Lake (Kırşehir), Turkey. *Gazi University Journal of Science* 18 (4): 563–567.
- ZAR J. H. 1996. *Biostatistical analysis*. Upper Saddle River, New Jersey: Prentice Hall, 662 p.